

Epilithic Soft Algae of Dilimi River in Jos, Nigeria

Cyril C. Ajuzie

Aquaculture, Freshwater and Marine Ecology Research Lab, Fisheries & Aquaculture Unit, Department of Animal Production, University of Jos, Nigeria
efulecy@yahoo.com

Abstract: River Dilimi flows through urban areas in Jos, Nigeria. As a result of this, a lot of human-generated pollutants find their way into the river. The locals attach a lot of socio-economic importance to the river. But the scientific community has shown minimal interest in the ecology of the river. Hence, there is a dearth of information in the literature about the biotas (especially soft algae) that inhabit the river. Epilithic soft algae were sampled from the river at two sites (an upstream site close to British-America bridge, and a downstream site at the pedestrian bridge, Unijos permanent site). Nutrients (N and P), biochemical oxygen demand (BOD), conductivity, and total dissolved solids (TDS) levels were relatively higher at the downstream site, which suffers more from anthropogenic pollution. Seven Divisions of soft algae were registered during this study. Cyanobacteria, Charophyta, Chlorophyta and Dinophyta were recorded at the upstream site. The fore-mentioned Divisions (excluding Dinophyta) plus Euglenophyta, Ochrophyta and Cryptophyta were observed in samples collected at the downstream site. Cyanobacteria was the most common group of soft algae at the upstream site with 82 % occurrence. At the downstream site, Chlorophyta was the most common group with 35 % occurrence, followed by Cyanobacteria (29 % occurrence) and Euglenophyta with 16 % occurrence. A total of 78 species of soft algae were recorded in this study. The downstream site was richer in species (57 species vs. 30 species at the upstream site), and had a higher diversity index value (3.89 vs. 2.67 Shannon index at the upstream site). The community similarity index between the two sites was low (11.5 %). This study is the first to describe the community of soft algae in River Dilimi, a grossly polluted river. Hence, the documented soft algae could be described as pollution tolerant organisms. [Ajuzie CC. **Epilithic Soft Algae of Dilimi River in Jos, Nigeria**. *Nat Sci* 2016;14(11):102-111]. ISSN 1545-0740 (print); ISSN 2375-7167 (online). <http://www.sciencepub.net/nature>. 16. doi: [10.7537/marsnsj141116.16](https://doi.org/10.7537/marsnsj141116.16).

Keywords: Dilimi River; epilithic soft algae; Nigeria; pollution tolerant organisms

1. Introduction

In lotic ecosystems, due to the main unidirectional flow of water, the first signs of eutrophication may be detected by changes in the periphytic community [1, 2]. And because the first signs of change often occur in attached communities [3, 4, 5], the biological monitoring of periphyton has been deemed a useful tool in the detection of anthropogenic impacts on rivers and streams. Occurrence and changes in the composition of periphytic species [6, 7, 8] are closely associated with environmental pollution [9]. Periphyton is, also, sensitive to the amount and type of pollutants. For example, species composition and abundance of periphyton have been reported to be highly dependent on the nitrogen/phosphorus (N:P) ratio [10].

Soft algae communities, as a constituent of periphyton, contribute immensely to the biodiversity associated with lotic ecosystems. They grow on pebbles, stones, boulders and bedrocks in rivers and other aquatic ecosystems. They form the basis of the aquatic food, and act as natural purification agents of freshwater bodies, since they absorb nutrients and other pollutants. They are very responsive to degradation of water quality (often changing in both taxonomic composition and biomass where even

slight contamination occurs). They can proliferate when high concentrations of nutrients occur in the water and velocities are low. They can provide habitat for many other organisms, especially rotifers [see 11, 12, 13]. They, thus, serve as micro environmental indicators of physical, chemical, and biological disturbances that occur in lotic ecosystems [see 12], and, hence, serve as indicators of biological integrity of freshwater bodies [e.g. 14, 15, 16, 17].

River Dilimi originates from the Jos Plateau and passes through five other Nigerian states (Jigawa, Kano, Yobe, Borno and Bauchi, where it is has acquired different names such as Hadejia, Jama'are, Kamadugu and Yobe River) before finally emptying into Lake Chad. The river supports several socio-economic activities of the locals [see 18, 19]. The people fish, wash clothes and bath in the river. They also harvest water from the river for domestic use, agriculture and block moulding. Farmers and vegetable-mongers wash harvested vegetables in the river before taking them to the market. However, irrespective of the diverse uses of the Dilimi River by the locals, the scientific community has paid little attention to the river [20]. Hence, the ecology, biology and taxonomy of biotas inhabiting the river are not

well studied. Soft algae of the river have, hitherto, not been studied and reported in the literature.

This work was designed primarily to take an inventory of epilithic soft algae in the River Dilimi, using two sampling sites that served only as reference

sites. Hence, samples collected for each site was pooled to form a composite sample. For this kind of study, many researchers [e.g. 21-23] suggest the collection of a single composite sample for each study location.

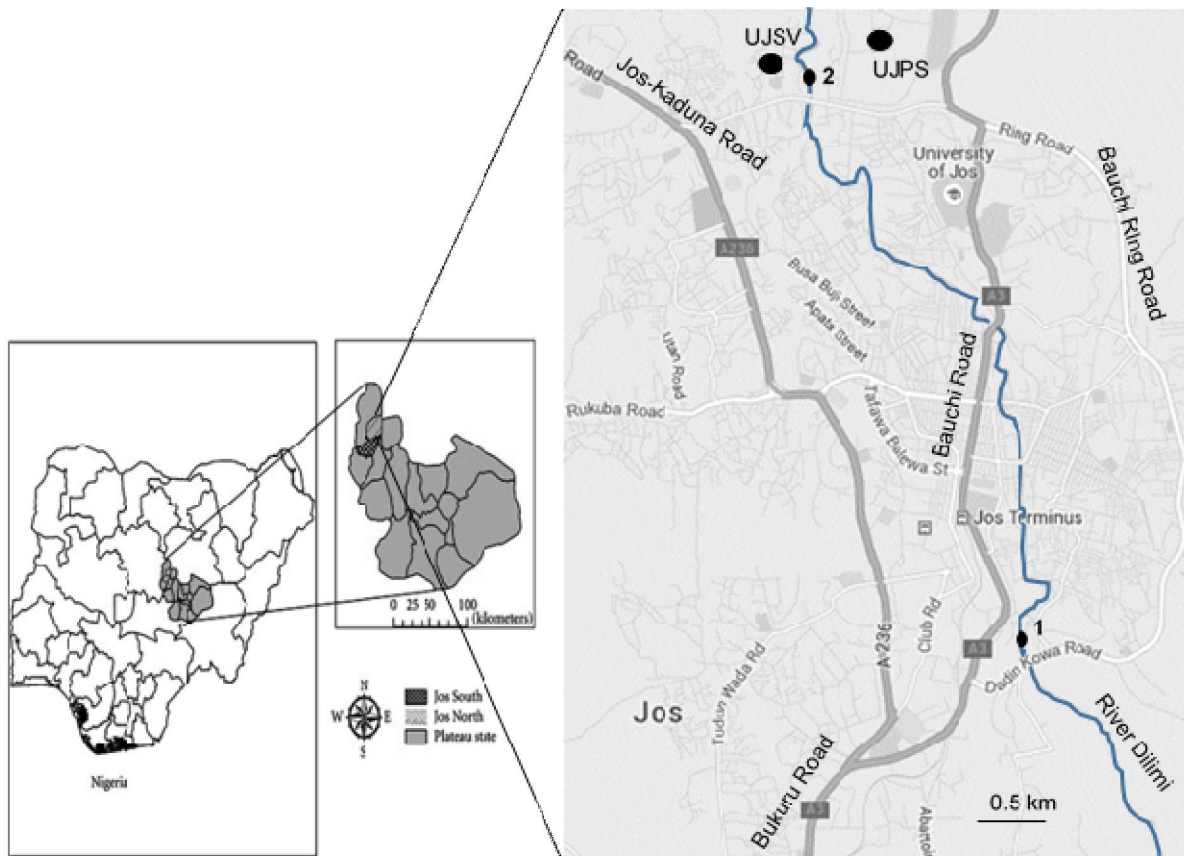


Figure 1. A section of Jos town showing River Dilimi and the study sites [1: upstream; 2: downstream; UJSV: University of Jos (Unijos) Students' Village; UJPS: Unijos Permanent Site]. Modified from Adebajo et al [68]

2. Material and Methods

2.1. Study Area

The urban section of Dilimi River runs through Jos North Local Government Area of Plateau State, Nigeria (Figure 1). Two sampling sites that included an upstream sampling station at about 200 m away from (and downstream of) the British-American bridge, and a downstream sampling location at the University of Jos students' pedestrian bridge, which links the University of Jos Students' Village (located on the left bank of the river) with the university's permanent site on the right bank. The area adjacent to the river banks at the British-American axis (because of the massive granite rocks that dot the area) have comparatively sparse human populations than the area adjacent to the river banks at the downstream axis (i.e. from ca. 400m after the British-American bridge to the students' pedestrian bridge at the permanent site of the University of Jos). By the time the river reaches

the permanent site of the University of Jos, it has passed through many densely populated poor neighbourhoods of Jos town, where houses and yards have direct link with the river, and the flood plains intensively farmed. A consequence of encroaching into the river banks was that in July 2012 the overflow of the river (after heavy rains) swept off many houses and farm lands that were situated along the banks [24]. Apart from these encroachments, the locals also defecate on the banks and in the river channel. Household organic and inorganic wastes, as well as wastes from business houses are ceaselessly emptied into the river by inhabitants of these poor neighbourhoods. The poor farming practices also enrich the river with nutrients and silt. The river water is, hence, discoloured throughout the year. In fact, the water has an odour.

2.2. Physico-chemical Parameters Studies

Temperature was measured on the spot using a mercury thermometer. Water conductivity, total dissolved solids (TDS) and pH were also measured on the spot with a multi-parameter water tester (HANNA® instruments). Nitrate nitrogen and phosphate phosphorus were equally measured on the spot with the JBL TESTSET™ reagents for iron, nitrate, and phosphates. Dissolved oxygen and biochemical oxygen demand (BOD₅) were determined by iodometric (Winkler) method [25]. Although I have already published the results elsewhere [see 20], they will still be shown in the present paper, because sampling for the two studies were carried out on the same day.

2.3. Collection of Epilithic Soft Algae

According to Stevenson [21], periphyton samples should be collected during periods of stable flow, since high flows can scour the substratum and result in flushing off the periphyton. Recovery after high discharge can be as rapid as seven days if severe scouring of the substrata did not occur [21]. Bearing this in mind, the two sites were sampled twice in May 2013. The first was before the first major rainfall of the year and the second was 10 days after the rainfall. Four submerged stones (one each from the riffles, runs, shallow pools and nearshore areas of the river) were sampled randomly at each sampling location, by wading into the river. Each stone was placed in a white laboratory tray. Soft algae were brushed off each of the stones, using a tooth brush and rinsed with limited quantity of river water. The cap of the sample holder (16.6 cm²) was used to define a sampling circle on each stone, by placing it on the stone. A circular mark was scratched on the stone around the outside of the cap with the tip of a scalpel blade [26]. Soft algae were sampled within the circle. The samples were preserved with 4 % formalin. This study was planned with emphasis on the spatial composition of the algae, with reference to the study sites. Thus, even though epilithic algae were sampled at two different times, samples from stones at each sampling station were pooled to form a composite sample for that location. The pooled samples were transferred to 250 ml sample bottles and distilled water added to bring the sample volume to 200 ml.

2.4. Identification and Enumeration of the Soft Algae

Each sample bottle was moderately shaken in order to get a homogenous solution before taking 50 µl subsample for microscopic analysis – i.e. identification and counting of the soft algae. The 50 µl subsample was dropped on a plane microscope slide and carefully covered with a cover slip to exclude bubbles. The slide was then transferred to the microscope stage for the analysis. Stancheva *et al.* [27] suggested the use of plane microscope slide

instead of a counting chamber for proper identification and counting of mixed microalgae species. Three hundred (300) soft algae units (cells or filaments) were identified to species level and enumerated at 400x magnification. Although larger counts may reduce uncertainties associated with organism counts [28], the benefit of increasing counts above 300 is not high [27]. The algae were viewed under randomly-selected six viewing fields, as suggested by Baffico *et al.* [29]. Several soft algae identification guides for freshwater ecosystems [including 26, 30-32] and the web were used in the identification of the species.

2.5. Species Density

Species density was calculated as follows: $C/A = (TN \times SV \times ACS) / (AVF \times NVF \times VSS \times SA)$, where C/A is the number of cells or filaments, as the case may be, per surface area of stone sampled; TN, total number of individuals; SV, sample volume; ACS, area of cover slip; AVF, area of viewing field at 400x magnification; NVF, number of viewing fields scanned; VSS, volume of subsample; and SA, surface area of stone sampled. The area of stone surface sampled was calculated as the surface area of an individual stone (mm²) multiplied by the total number of stones sampled for that site [see 26].

2.6. Percent (%) Composition of Soft Algae Species

This was calculated for each species by dividing species density (C/A) of each species by the total density summed from values recorded for each of the species each site, and the result multiplied by 100. For example, % Composition of a species “A” was given as: $A = (a/b) \times 100 \%$, where: a is the calculated C/A for the species A, and b is $\sum C/A$ for a sampled location [e.g. 21].

2.7. Species diversity

Shannon Index (H') was used to calculate the species diversity index at each study site. This was calculated thus: $H' = - \sum [(ni/N) \times \ln (ni/N)]$, where: ni = number of individuals of each species (the i^{th} species), N = total number of individuals for the site, and \ln = the natural log of the number.

2.8. Community Similarity index (%)

The similarity index (%) of soft algae between the two sites was obtained by multiplying a calculated Jaccard index by 100. Jaccard Index (J) was calculated thus: $J = sc / (sa + sb + sc)$, where: sa and sb are the numbers of species unique to samples a and b, respectively, and sc is the number of species common to the two samples.

2.9. Student's t-Test

Paired Two Sample for Means t-Test [$P(T \leq t)$ two-tail] was performed to further test if differences observed in some of the data sets were statistically significant ($\alpha = 0.05$).

3. Results

3.1. Physico-chemical Parameters Studies

Water temperature was lower than air temperature, but both depended on both cloud cover and time of the day (increasing as the sun rises on a cloudless sky). Temperatures were lower at the upstream site (23.8 ± 3.9 °C) than at the downstream site (27.8 ± 2.6 °C). Dissolved oxygen concentration was higher upstream (7.1 ± 0.07 mg l⁻¹) than downstream (4.3 ± 0.28 mg l⁻¹), but the difference was not statistically significant. Also, there was no statistically significant difference in TDS, Fe, NO₃ and PO₄ concentrations between the two sites. N:P ratio was lower downstream (0.8), and higher upstream (30). The river is weakly alkaline as observed from the pH readings. The difference in biochemical oxygen demand between the two sites was statistically significant. So, too, was the difference in conductivity levels (Table 1).

3.2. The soft algae

Four Divisions of soft algae (Cyanobacteria, Dinophyta, Charophyta and Chlorophyta) were

recorded at the upstream site, and six (Cyanobacteria, Charophyta, Chlorophyta, Euglenophyta, Ochrophyta and Cryptophyta) at the downstream site (Tables 2 and 3). Cyanobacteria was the most common group among the soft algae community at the upstream site (Figure 2). At the downstream site Chlorophyta was the most common Division (Figure 3). Dinoflagellates, cryptophytes and ochrophytes were rare. Among the Cyanobacteria the genus *Gloeocapsa* was the most common of all the genera recorded with 45.23 % occurrence at the upstream site. Within the Division Euglenophyta, *Euglena* (11.16 %) was the most common genus with *E. viridis* (9.88 %) the most common species (Table 3). There was a total of 174,582 units of soft algae per square mm of stone surface at the upstream site, and 210,290 units mm⁻² at the downstream site. Species richness was higher at the downstream site than at the upstream site. A similar observation was made for species diversity (Shannon index). Community similarity index between the two sites was 11.5 % (Table 4).

Table 1. Physical and chemical properties of the sampling sites

Parameter	Upstream	Downstream
Air Temperature (°C)	26.3 ± 3.9	30.8 ± 2.6
Water Temperature (°C)	23.8 ± 3.4	27.8 ± 1.5
Dissolved Oxygen (mg l ⁻¹)	7.1 ± 0.07	$4.3^{ns1} \pm 0.28$
BOD ₅	$0.75^{*1} \pm 0.07$	3.7 ± 0.14
Fe (mg l ⁻¹)	$0.05^{ns2} \pm 0$	0.65 ± 0.07
NO ₃ (mg l ⁻¹)	$0.9^{ns3} \pm 0.14$	1.06 ± 0.08
PO ₄ (mg l ⁻¹)	$0.03^{ns4} \pm 0.14$	1.3 ± 0.7
NO ₃ :PO ₄ ratio	30	0.8
pH	7.8 ± 0.3	7.9 ± 0.4
Conductivity (μs cm ⁻¹)	$213^{*2} \pm 1.41$	512 ± 51.62
Total Dissolved Solids (ppm)	$112^{ns5} \pm 7.07$	257 ± 26.16

N/B: ns = not statistically significant; * = statistically significant; ns¹ paired t(1) = 11.4, p = 0.056; ns² paired t(1) = 12, p = 0.053; ns³ paired t(1) = 1, p = 0.50; ns⁴ paired t(1) = 2.49, p = 0.24; ns⁵ paired t(1) = 10.7, p = 0.059 *¹ paired t(1) = 19.67, p = 0.032; *² paired t(1) = 19.67, p = 0.03

4. Discussion

4.1. Physical and Chemical Parameters

Water temperature was lower than that of air, and increased as the sampling time approached noon. Data and sample collections were carried out on the same day, beginning from the upstream site, between 09:00 and 12:00 hours. The foregoing explains why temperature was relatively higher at the downstream site. Although the difference in dissolved oxygen concentration at the two sites was not statistically significant, the higher mean BOD value at the downstream site corroborates the findings associated with nutrient concentration at the two sites. The downstream site is subjected to more nutrient loads than the upstream site. The increased nutrient load

downstream (especially of organic pollutants) has the potential to cause an increase in bacteria load. An increase in bacteria load would lead to an increase in bacterial activity, which will, in turn, quicken the consumption of dissolved oxygen [see 33]. The major source of N and P loadings in the downstream section of the study site is untreated sewage from homes, business centres, and direct defecation on the river banks and in river channel. Phosphorus enrichment, for example, is associated with increased microbial biomass and activity, resulting in faster rates of decomposition and nutrient cycling downstream of aquatic ecosystems [e.g. 34]. Jarvie *et al.* [35] observed that phosphorus treatment at selected major sewage treatment works in the upper Thames basin in

the UK resulted in significant reductions in in-stream P concentrations. There is no such treatment plant associated with the Dilimi River. The findings in this study are in line with the observation that nutrient enrichments are major water quality concerns in lotic ecosystems [see 36-38]. Soil tillage and fertilizer applications are also common practices along the

downstream axis of the study site. These habits indirectly load soil materials and nutrients into the river, via runoff. Control measures for runoff loading of both N and P would include containment and treatment of manure, decreased use of fertilizers, and a control of soil tillage practices [see 37].

Table 2. Density and % composition of epilithic cyanobacteria, euglenophytes, and charophytes at the study sites

Taxon	Density Upstream Units/mm² (%)	Density Downstream Units/mm² (%)
Cyanobacteria		
<i>Aphanizomenon flos-aquae</i> Ralfs ex Bornet & Flahault	2934 (1.68)	5624 (2.67)
<i>Chamaesiphon incrustans</i> Grunow	3179 (1.82)	
<i>Coelosphaerium aponina</i> Kützing	3179 (1.82)	
<i>Coelosphaerium naegelianum</i> Unger		2690 (1.28)
<i>Gloeocapsa punctata</i> Nägeli		4890 (2.33)
<i>Gloeocapsa rupestris</i> Kützing	51345 (29.41)	
<i>Gloeocapsa sanguinea</i> (C. Agardh) Kützing	3179 (1.82)	
<i>Gloeocapsa turgida</i> (Kützing) Hollerbach	24450 (14.00)	4646 (2.21)
<i>Hyella fontana</i> Huber & Jadin		3668 (1.74)
<i>Lyngbya major</i> Meneghini ex Gomont		6113 (2.91)
<i>Lyngbya martensiana</i> Meneghini ex Gomont	4157 (2.38)	
<i>Microcoleus lacustris</i> Farlow ex Gomont	2934 (1.68)	
<i>Microcystis aeruginosa</i> (Kützing) Kützing	3179 (1.82)	
<i>Nodularia spumigena</i> Mertens ex Bornet & Flahault	19560 (11.20)	
<i>Nostoc entophyllum</i> Bornet & Flahault		2934 (1.40)
<i>Oscillatoria jasorvensis</i> Vouk	4890 (2.80)	4646 (2.21)
<i>Oscillatoria limnetica</i> Lemmermann	2934 (1.70)	
<i>Oscillatoria platensis</i> (Gomont) Bourrelly	3668 (2.10)	2690 (1.28)
<i>Oscillatoria simplicissima</i> Gomont		3179 (1.51)
<i>Plectonema tomasianum</i> Bornet ex Gomont		3668 (1.74)
<i>Pseudanabaena minuta</i> Skuja		2690 (1.28)
<i>Rivularia biasolettiana</i> Meneghini ex Bornet & Flahault	6113 (3.50)	2445 (1.16)
<i>Stigonema mamillosum</i> C. Agardh ex Bornet & Flahault		7336 (3.49)
<i>Synechococcus leopoliensis</i> (Raciborski) Komárek		3912 (1.86)
<i>Synechocystis aquatilis</i> Sauvageau	3668 (2.10)	
Charophyta		
<i>Closterium aciculare</i> T. West		2934 (1.40)
<i>Closterium ehrenbergii</i> Meneghini ex Ralfs		3179 (1.51)
<i>Cosmarium candianum</i> Delponte	734 (0.42)	
<i>Cosmarium circulare</i> Reinsch	1956 (1.12)	2934 (1.40)
<i>Cosmarium cucurbita</i> Brébisson ex Ralfs		1467 (0.70)
<i>Cosmarium margaritifera</i> Meneghini ex Ralfs	2690 (1.54)	
<i>Cosmarium praemorsum</i> Brébisson		2690 (1.54)
<i>Cylindrocystis brebissonii</i> (Ralfs) De Bary		2934 (1.36)
<i>Euastrum montanum</i> West & G.S. West		734 (0.42)
<i>Gonatozygon monotaenium</i> De Bary		2690 (1.28)
<i>Mougeotia floridana</i> Transeau		4157 (1.98)
<i>Netrium digitus</i> (Brébisson ex Ralfs) Itzigsohn & Rothe		2690 (1.28)
<i>Penium polymorphum</i> (Perty) Perty		3423 (1.63)
<i>Spirogyra gracilis</i> Kützing	2690 (1.54)	2690 (1.28)
<i>Staurastrum anatinum</i> Cooke & Wills		978 (0.47)
<i>Staurastrum cingulum</i> (West & G.S. West) G.M. Smith		1956 (0.93)
<i>Staurastrum paradoxum</i> Meyen ex Ralfs	1223 (0.70)	
<i>Zygnema stellinum</i> (O.F. Müller) C. Agardh	1712 (0.98)	
<i>Zygogonium ericetorum</i> Kützing	8313 (4.76)	

Table 3. Density and % composition of epilithic chlorophytes, dinophytes, cryptophytes and ochrophytes at the study sites

Taxon	Density Upstream Units/mm ² (%)	Density Downstream Units/mm ² (%)
Chlorophyta		
<i>Botryococcus braunii</i> Kützing		3423 (1.63)
<i>Bulbochaete</i> sp.		6113 (2.90)
<i>Chlamydomonas acidophila</i> Negoro		2690 (1.28)
<i>Chlamydomonas stellata</i> O. Dill		2690 (1.28)
<i>Chlamydomonas</i> sp.		2690 (1.28)
<i>Chlorella vulgaris</i> Beyerinck [Beijerinck]	2690 (1.54)	2690 (1.28)
<i>Cladophora</i> sp.		3668 (1.74)
<i>Coelastrum cambricum</i> W. Archer		2934 (1.40)
<i>Coelastrum microporum</i> Nägeli	978 (0.56)	
<i>Eremosphaera viridis</i> De Bary		3423 (1.63)
<i>Eudorina elegans</i> Chodat		2690 (1.28)
<i>Geminella minor</i> (Nägeli) Heering		3668 (1.74)
<i>Gongrosira incrustans</i> (Reinsch) Schmidle		1223 (0.58)
<i>Monoraphidium</i> sp.		2690 (1.28)
<i>Oedogonium acrosporum</i> De Bary ex Hirn	2934 (1.68)	
<i>Oedogonium anomalum</i> Hirn		2690 (1.28)
<i>Oedogonium subellipsoideum</i> Tiffany		2690 (1.28)
<i>Oocystis lacustris</i> Chodat	2690 (1.54)	3179 (1.51)
<i>Pediastrum tetras</i> (Ehrenberg) Ralfs		3423 (1.63)
<i>Quadrigula pfitzeri</i> (Schröder) G.M.Smith		3668 (1.74)
<i>Scenedesmus arcuatus</i> (Lemmermann) Lemmermann		4401 (2.09)
<i>Schizomeris leibleinii</i> Kützing		2690 (1.28)
<i>Sphaerocystis schroeteri</i> Chodat		4890 (2.32)
<i>Stigeoclonium aestivale</i> (Hazen) Collins		7336 (3.49)
<i>Trentepohlia umbrina</i> (Kützing) Bornet	734 (0.42)	
Cryptophyta		
<i>Cryptomonas ovata</i> Ehrenberg		2690 (1.28)
Dinophyta		
<i>Gymnodinium rotundatum</i> Klebs	2445 (1.40)	
Euglenophyta		
<i>Colacium cyclopicola</i> (J.Gicklhorn) Woronichin & Popova		2445 (1.16)
<i>Euglena viridis</i> (O.F.Müller) Ehrenberg		20783 (9.88)
<i>Euglena anabaena</i> Mainx		2690 (1.28)
<i>Phacus unguis</i> Pochmann		3668 (1.74)
<i>Trachelomonas</i> sp.		3912 (1.86)
Ochrophyta		
<i>Ophiocytium cochleare</i> (Eichwald) A.Braun		2690 (1.28)
<i>Vacuolaria virescens</i> Cienkowski		2690 (1.28)

Though many researchers are of the opinion that P is the major pollutant that constrains algae production in freshwater ecosystems [see 37, 39], a comparison of results of algal bioassays and nutrient concentrations in freshwater bodies suggests that an N:P ratio above 17 indicates P limitation, a ratio below 10 indicates N limitation and values between 10 and 17 indicate that either of the nutrients may be limiting [see 34, 40-44]. From the foregoing it could

be stated that at the study sites, P is the limiting nutrient upstream, and N downstream.

While the differences in both NO₃ and PO₄ concentrations at the upstream and downstream sites were not statistically significant, the concentrations recorded for these compounds during this study (and the filthiness of the immediate surroundings of the river) suggest that the section of the river studied is actively polluted. It has been argued that nitrate-nitrogen concentrations above 3 mg l⁻¹ and any

detectable amounts of total phosphorus (usually above 0.025 mg l⁻¹) may be indicative of pollution from fertilizers, manures or other nutrient-rich wastes [see 45]. The downstream site is affected by a heavily populated and largely poor neighbourhood with a very

poor sanitary habit. The amount of municipal wastes and raw sewage from these settlements that find their way into Dilimi River (though yet to be quantified and reported in the literature) is disturbing [20].

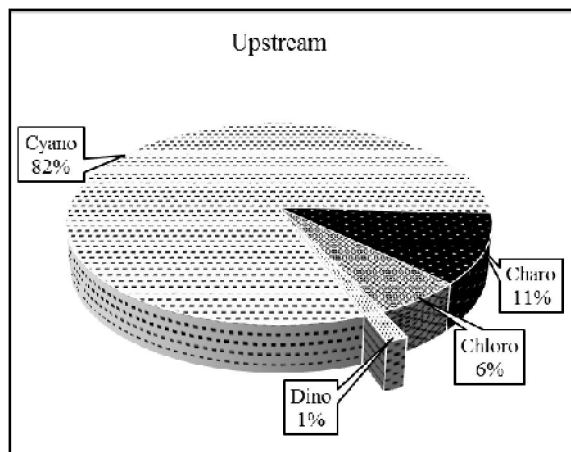


Figure 2. Composition (%) of soft algae upstream (Cyanobacteria 82, Charophyta 11, Chlorophyta 6, & Dinophyta 1)

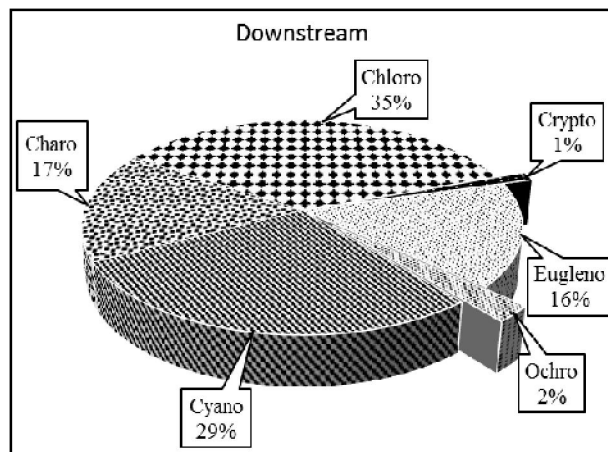


Figure 3. Composition (%) of soft algae downstream (Chlorophyta 35, Cyanobacteria 29, Charophyta 17, Euglenophyta 16, Ochrophyta 2, & Cryptophyta 1)

Table 4. Species richness, diversity and % community similarity indices of epilithic soft algae at the sampling sites

Index	Upstream	Downstream
Species Richness	30 species	57 species
Shannon (H ²)	2.67	3.89
Community Similarity (%)	11.5	

The river is weakly alkaline and, thus, (in the absence of pollution) has the capacity to support many forms of aquatic life. The large number of species of soft algae witnessed during this study supports this assertion. This is in contrast to acidic freshwater bodies, which are characterized by benthic algal communities with low diversity [29, 46]. The high electrical conductivity and TDS values witnessed at the downstream site are indications that this section of the river had more solutes (including chemical ions) than the upstream site. Human activities greatly impact the conductivity and concentration of TDS in lotic ecosystems. As earlier noted, the downstream section of the river is heavily loaded with domestic and industrial wastewater, and untreated sewage, as well as suffers impact from poor farming practices on the floodplains. Most probably, a high bacterial activity (mineralisation of organic wastes) must have played a role in the elevation of EC and TDS at the downstream site.

4.2. The Soft Algae

Hitherto, there is no information in the literature on soft algae in Dilimi River, Jos, Nigeria. This observation is not unique to the river. For example,

Potapova [47] reported that the taxonomy and ecology of many riverine algae in North America have yet to be studied, just as Porter [48] observed that the autecology of soft algae is poorly understood. Many researchers in Nigeria work on planktonic algae [e.g. 49, 50]. Only a few [e.g. 51] work on attached algae in lotic ecosystems. Although Tiseer et al. [51] sampled phytoplankton and attached algae in the Samaru Stream in Zaria, Nigeria, their report failed to show which algae were planktonic and which were periphytic. This made it difficult for any comparison to be made between periphytic soft algae species in Samaru Stream and those in Dilimi River.

The downstream site of the Dilimi River was richer in species and had a higher species diversity index than the upstream site. The comparatively higher nutrient loads (pollutants) at the downstream axis of the river must have contributed to these findings. Pearson and Rosenberg [33] and Krewer and Holm [52] are of the opinion that if pollutants are readily available as food for algae, they will easily bring about an increase in population via bio-stimulation. And results from several bioassay techniques have demonstrated benthic algal growth

stimulation with additions of P [52, 53], N [54], and both P and N [55, 56].

The soft algae in the study sites could be referred to as pollution-tolerant algae (see 15, 33, 47, 57-59). Although species of *Cladophora* do have contrasting ecological preferences [26], the genus is often associated with eutrophication [see 60]. Similarly, species of *Closterium* [57, 61], *Chlorella* [47, 57, 62, 63], *Cosmarium* [57], *Oedogonium* [47], *Oscillatoria*, [64], *Scenedesmus* [13, 62, 65, 66], *Stigeoclonium*, [13, 59], and *Euglena* [67] have been cited as indicators of polluted waters.

5. Conclusion

The present study succeeded in qualifying and quantifying epilithic soft algae in the Dilimi River. The study also presented some autecological information about the soft algae. In view of the fact that the section of the River Dilimi studied is nutrient-enriched, the soft algae recorded in this study could be described as pollution tolerant algae.

Acknowledgements

This work was supported by the tetfund of Nigeria research grant received in 2013. Mr. A. Uja determined the DO and BOD of water samples taken from the study sites.

References

- Soballe DM, Kimmel BL. A largescale comparison of factors influencing phytoplankton abundance in rivers, lakes, and impoundments. *Ecology* 1987; 68: 1943-1954.
- Stevenson RJ, Stoermer EF. Abundance patterns of diatoms on *Cladophora* in Lake Huron with respect to a point source of wastewater treatment plant effluent. *Journal of Great Lakes Research* 1982; 8: 184-195.
- Wetzel RG, Hough RA. Productivity and role of aquatic macrophytes in lakes. An assessment. *Polskie Archiwum Hydrobiologii* 1973; 20: 9-19.
- Heinonen I, Herve S. A rapid biological method for the monitoring of eutrophication. *Archiv für Hydrobiologie* 1984; 101: 135-42.
- Descy JI, Micha JC. Use of biological indices of water quality. *Statistical Journal of the United Nations Economic Commission for Europe* 1988; 5: 249-261.
- Economou-Amilli A. Periphyton analysis for the evaluation of water quality in running waters of Greece. *Hydrobiologia* 1980; 74: 39-48.
- Marcus MD. Periphytic community response to chronic nutrient enrichment by a reservoir discharge. *Ecology* 1980; 61: 387-99.
- Rushforth SR, Brock JT. Attached diatom communities from the lower Truckee River, summer and fall, 1986. *Hydrobiologia* 1991; 224: 49-64.
- Baffico GD, Pedrozo FL. Growth factors controlling periphyton production in a temperate reservoir in Patagonia used for fish farming. *Lakes and Reservoirs: Research and Management* 1996; 2: 243-249.
- Tilman D, Kiesling R, Sterner R, Kilham SS, Johnson EA. Green, bluegreen and diatom algae: Taxonomic differences in competitive ability for phosphorus, silicon and nitrogen. *Archiv für Hydrobiologie* 1986; 106: 473-85.
- Stevenson RJ. Patterns of benthic algae in aquatic ecosystems. In: Stevenson RJ, Bothwell MB, Lowe RL, eds. *Algal Ecology: Freshwater Benthic Ecosystems*. 1996; pp. 3-26, Academic Press, San Diego, CA.
- Blinn DW, Herbst DB. Use of diatoms and soft algae as indicators of environmental determinants in the Lahontan Basin, USA. *Annual Report for California State Water Resources Board, California, USA*. 2003.
- Wehr JD, Sheath RG. Freshwater habitats of algae. In: Wehr JD, Sheath RG, eds. *Freshwater Algae of North America*. 2003; pp 11-57, Academic Press, New York.
- Palmer CM. Algae in Water Supplies of Ohio. *Ohio Journal of Science* 1962; 62: 225-244.
- Palmer CM. A composite rating of algae tolerating organic pollution. *Journal of Phycology* 1969; 5: 78-82.
- Fjordingstad E. Some remarks on a new saprobic system. In: Tarzwell CM, ed. *Third Seminar on Biological Problems in Water Pollution*. 1965; 232-235 pp. U.S. Public Health Service, Cincinnati, OH.
- Hill BH, Herlihy AT, Kaufmann PR, Stevenson RJ, McCormick FH, Johnson CB. Use of epiphyton assemblage data as an index of biotic integrity. *Journal of North American Benthological Society* 2000; 19: 50-67.
- Abari Y. On the optimal use of Hadejia-Jama'are-Kamadugu-Yobe River Basin. 2009. <http://newstoweronline.blogspot.com.ng/2009/09/on-optimal-use-of-hadejia-jamaare-27.html> (Accessed on 22 January 2016).
- Idegu YA, Kwapyel E. Plateau leads save Lake Chad battle. 2014. <http://thenationonline.net/plateau-leads-save-lake-chad-battle/> (Accessed on 22 January 2016).
- Ajuzie CC. Epilithic diatoms of urban River Dilimi, Jos, Nigeria. *Nature and Science* 2016; 14(8): 88-97.
- Stevenson RJ. Benthic algal community dynamics in a stream during and after a spate. *Journal of North American Benthological Society* 1990; 9: 277-88.
- Kelly MG, Cazaubon A, Coring E, Dell'uomo A, Ector L, Goldsmith B, Guasch H, Hürlimann J,

- Jarlman A, Kawecka B, Kwandrans J, Laugaste R., Lindstrøm E-A, Leitao M, Marvan P, Padisák J, Pipp E, Prygiel J, Rott E, Sabater S, Van Dam H, Vizinet J. Recommendations for the routine sampling of diatoms for water quality assessments in Europe. *Journal of Applied Phycology* 1998; 10: 215–224.
23. Fetscher AE, Stancheva R, Kociolok JP, Sheath RG, Stein ED, Mazor RD, Ode PR, Busse LB. Development and comparison of stream indices of biotic integrity using diatoms vs. non-diatom algae vs. a combination. *Journal of Applied Phycology* 2014; 26: 433–450.
 24. Ezema O. Flood menace in Nigeria: Plateau State flood disaster. NICE's Events and News. The Nigerian Institution of Civil Engineers. 2013. <http://www.nice-nigeria.org/component/jumi/news-in-full?id=123#sthash.RD0gcohM.dpf> (Accessed on 22 January 2016).
 25. USGS. National Field Manual for the Collection of Water-Quality Data: Techniques of Water-Resources Investigations Book 9. Handbooks for Water-Resources Investigations. U.S. Department of the Interior and U.S. Geological Survey. 2015.
 26. Biggs BJF, Kilroy C. Stream periphyton monitoring manual. Published by NIWA for Ministry for the Environment, New Zealand. 2000.
 27. Stancheva R, Fetscher AE, Sheath RG. A novel quantification method for stream-inhabiting, non-diatom benthic algae, and its application in bioassessment. *Hydrobiologia* 2012; 684: 225–239.
 28. Birks HJB. Numerical methods for the analysis of diatom assemblage data. In: Smol JP, Stoermer EF, eds. *The Diatoms: applications for the environmental and earth sciences*, 2nd ed. 2010; pp. 23–57, Cambridge University Press, Cambridge, MA.
 29. Baffico GD, Diaz MM, Wenzel MT, Koschorreck M, Schimmele M, Neu TR, Pedrozo F. community structure and photosynthetic activity of epilithon from a highly acidic (pH < 2) mountain stream in Patagonia, Argentina. *Extremophiles* 2004; 8: 463–473.
 30. Bourrelly P. *Les Algues d'Eau Douce*. Tome I: *Les Algues Vertes*. 1966; 511 pp, Editions N. Boubée et Cie., Paris.
 31. Durand JR, Leveque C. *Flore et faune aquatiques de l'Afrique Sahelo-Soudanienne*. 2 Vol. Ed. de l'Office de la Recherche Scientifique et Technique Outre-Mer. Doc. Techn. No. 44 Paris. 1980.
 32. Pentecost A. *Introduction to Freshwater Algae*. 1984; 247 pp., Kingprint Limited, Richmond, Surrey.
 33. Pearson TH, Rosenberg R. Macrobenthic succession in relation to organic enrichment and pollution of the environment. *Oceanography and Marine Biology Annual Review* 1978; 16: 229–311.
 34. McCormick PV, Newman S, Miao SL, Reddy KR., Rawlik D, Fitz HC, Fontaine TD, Marley D. Ecological Needs of the Everglades. In: Redfield G, ed. *Everglades Interim Report*. 1998; 3.1–3.63 pp, South Florida Water Management District, West Palm Beach, Florida.
 35. Jarvie HP, Lycett E, Neal C, Love A. Patterns in nutrient concentrations and biological quality indices across the upper Thames river basin, UK. *Science of the Total Environment* 2002; 282–283: 263–294.
 36. Millennium Ecosystem Assessment. *Ecosystems and human well-being: Status and trends*. Cambridge Univ. Press, Cambridge, UK. 2005.
 37. Carpenter SR. Phosphorus control is critical to mitigating eutrophication. *PNAS* 2008; 105(32): 1039–11040.
 38. Porter SD, Mueller DK, Spahr NE, Mark D, Munn MD, Dubrovsky NM. Efficacy of algal metrics for assessing nutrient and organic enrichment in flowing waters. *Freshwater Biology* 2008; 53: 1036–1054.
 39. Schindler D W. Evolution of phosphorus limitation in lakes. *Science* 1977; 195: 260–262.
 40. Ulén B. Seston and sediment in Lake Norrviken. I. Seston composition and sedimentation. *Schweizerische Zeitschrift für Hydrologie* 1978; 40: 262–286.
 41. Forsberg C, Ryding SO. Eutrophication parameters and trophic state indices in 30 Swedish waste-receiving lakes. *Archiv für Hydrobiologie* 1980; 89: 189–207.
 42. Fu J, Winchester JW. Inference of nitrogen cycling in three watersheds of northern Florida, USA, by multivariate statistical analysis. *Geochimica et Cosmochimica Acta* 1994; 58 (6): 1591–1600.
 43. Hellström T. An empirical study of nitrogen dynamics in lakes. *Water Environment Research* 1996; 68: 55–65.
 44. Ekholm P. N: P ratios in estimating nutrient limitation in aquatic systems. Finnish Environment Institute. 2008. http://www.cost869.alterra.nl/fs/fs_npratio.pdf (Accessed on 06 February 2016).
 45. POND FACTS #2. Interpreting Water Tests for Ponds and Lakes. Pond Fact #2 Series. <http://extension.psu.edu/natural-resources/water/drinking-water/water-testing/testing/interpreting-water-tests-for-ponds-and-lakes/> (Accessed on 22 January 2016).
 46. Whitton B, Diaz B. Influence of environmental factors on photosynthetic species composition in highly acidic waters. *Verh Int Verein Limnol* 1981; 21: 1459–1465.
 47. Potapova MG. Relationships of soft-bodied algae to water-quality and habitat characteristics in U.S. Rivers. Analysis of the NAWQA national data set: Academy of Natural Sciences of Philadelphia, Patrick Center Report 05-08. 2005. <http://diatom.acnatsci.org/autecology/> (Accessed on 06 February 2016).

48. Porter SD. Algal attributes: An autecological classification of algal taxa collected by the National Water Quality Assessment Program. US Geological Survey Data Series 329. 2008. <http://pubs.usgs.gov/ds/ds329/> (Accessed on 06 February 2016).
49. Yakubu AF, Sikoki FD, Horsefall JRM. An investigation into the physicochemical conditions and planktonic organisms of the lower reaches of the Nun River. *Journal of Applied Sciences and Environmental Management* 1998; 1: 46-51.
50. Olaleye, VF, Adedeji AA. Water and planktonic quality of a palm oil effluent impacted river in Ondo State, Nigeria. *International Journal of Zoological Research* 2005; 1: 15-20.
51. Tiseer FA, Tanimu Y, Chia AM. Seasonal occurrence of algae and physicochemical parameters of Samaru Stream, Zaria, Nigeria. *Asian Journal of Earth Sciences* 2008; 1: 31-37.
52. Krewer J A and Holm H W. The phosphorus-chlorophyll a relationship in periphytic communities in a controlled ecosystem. *Hydrobiologia* 1982; 94: 173-176.
53. Fairchild GW, Lowe RL. Artificial substrates which release nutrients: effects on periphyton and invertebrate succession. *Hydrobiologia* 1984; 114: 29-37.
54. Grimm NB, Fisher SG. Nitrogen limitation in a Sonoran Desert stream. *Journal of North American Benthological Society* 1986; 5: 2-15.
55. Stockner JG, Shortreed KRS. Enhancement of autotrophic production by nutrient addition in a coastal rainforest stream on Vancouver Island. *Journal of the Fisheries Research Board of Canada* 1978; 35: 28-34.
56. Pringle CM, Bowers JA. An in situ substratum fertilization technique: diatom colonization on nutrient enriched substrata. *Canadian Journal of Fisheries and Aquatic Sciences* 1984; 41: 1247-1251.
57. Sladeczek V. System of water-quality from the biological point of view. *Archiv für Hydrobiologie, Beiheft 7, Ergebnisse der Limnologie, Heft 1973; 7: 1-218.*
58. Fairchild GW. Algal periphyton growth on nutrient-diffusing substrates: An in situ bioassay. *Ecology* 1985; 66 (2): 465-472.
59. Carrick HJ, Lowe RL. Response of Lake Michigan benthic algae to in situ enrichment with Si, N, and P. *Canadian Journal of Fisheries and Aquatic Sciences* 1988; 45: 271-279.
60. Biggs B. Patterns in benthic algae of streams. In: Stevenson RJ, Bothwell ML, Lowe RL, eds. *Algal Ecology. Freshwater Benthic Ecosystems*. 1996; pp. 31-56, Academic Press, New York.
61. Komárek J, Anagnostidis K. Cyanoprokaryota. 2. Oscillatoriales. In: Büdel B, Gärtner G, Krienitz L, Schagerl M, eds. *Susswassseflora von Mitteleuropa 19/2*. Elsevier GmbH, München. 2005.
63. Palmer CM. Algae in water supplies. US Department of Health, Education and Welfare, Public Health Service, Cincinnati. (Public Health Service publication No. 657). 1959.
64. Palmer CM. Algae. In: Parrish FK, ed. *Keys to Water-Quality Indicative Organisms of the Southeastern United States*. U.S. EPA, Environmental Monitoring and Support Laboratory, Office of Research and Development, Cincinnati, OH. 1975.
65. Van Landingam SL. Guide to identification, environmental requirements and pollution tolerance of freshwater blue-green algae (Cyanophyta). U.S. EPA, Environmental monitoring and support laboratory, Office of research and development, Cincinnati, OH. 1982.
66. Prescott GW. Algae of the Western Great Lakes area. Cranbook Institute of Science, Bloomfield Hills, MI. 1951.
67. Taylor WD, Williams LR, Hern SC, Lambou VW, Howard CL, Morris FA, Morris MK. Phytoplankton water-quality relationships in U.S. lakes, Part VIII: algae associated with or responsible for water-quality problems. U.S. EPA, Environmental Monitoring Systems Laboratory, Las Vegas, NV. 1981.
68. Raut KS, Kachare SV, Pathan TS, Shinde SE, Dabhade VF, Sonawane DL. Utilization of algae as pollution indicators of water quality at Nagapur and Chandapur Dams near Parli. V. Town Dist. Beed Maharashtra, India. *International Journal of Current Research* 2010; 4: 052-054.
69. Adebajo MC, Ademola SI, Oluwaseun A. (2012). Seroprevalence of fowl pox antibody in indigenous chickens in Jos North and South Council Areas of Plateau State, Nigeria: Implication for vector vaccine. *ISRN Vet. Sci.* 2012. doi: 10.5402/2012/154971.

10/2/2016